The Carbon Footprint of Water

by

Bevan Griffiths-Satenspiel and Wendy Wilson

A River Network Report
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Executive Summary

The decisions being made today regarding the management of water and energy resources will profoundly affect our economic and environmental future. A large amount of energy is expended to supply, treat and use water. This report explores the energy and carbon emissions embedded in the nation's water supplies. We have developed a baseline estimate of water-related energy use in the United States, as well as a comparative overview of the energy embedded in different water supplies and end-uses. We include numerous examples of how water management strategies can protect our freshwater resources while reducing energy and carbon emissions. This information is intended to help river and watershed groups, policy makers and water managers understand the magnitude of water-related energy use and evaluate the potential to reduce carbon emissions through water conservation, efficiency, reuse and low impact development strategies.

Through our analysis of primary and secondary research, we estimate that U.S. water-related energy use is at least 521 million MWh a year—equivalent to 13% of the nation's electricity consumption. While this appears to be a conservative estimate of water-related energy use, our findings suggest that the carbon footprint currently associated with moving, treating and heating water in the U.S. is at least 290 million metric tons a year. The CO2 embedded in the nation’s water represents 5% of all U.S. carbon emissions and is equivalent to the emissions of over 62 coal fired power plants.¹

Most significantly, the carbon footprint of our water use is likely growing for several reasons. Climate change is predicted to have numerous adverse affects on freshwater resources, rendering many available water supplies far less reliable. With water demand growing and many local, low-energy supplies already tapped, water providers are increasingly looking to more remote or alternative water sources that often carry a far greater energy and carbon cost than existing supplies. Furthermore, the adoption of higher water treatment standards at the state and federal levels will increase the energy and carbon costs of treating our water and wastewater.
Water conservation, efficiency, reuse and Low Impact Development (LID) strategies should be targeted to achieve energy and greenhouse gas emissions reductions. Research from the California Energy Commission suggests that programs focusing on these kinds of water management strategies can achieve energy savings comparable to traditional energy conservation measures at almost half the cost. Water management policies that promote water conservation, efficiency, reuse and low impact development can reduce energy demand and substantially decrease carbon emissions. The total energy savings potential of these strategies has yet to be assessed. However, numerous case studies illustrate the effectiveness of saving energy with water-based approaches. A few examples of these savings include:

- Retrofitting water using fixtures and appliances reduces hot water use by approximately 20%. If every household in the United States installed efficient fixtures and appliances, residential hot water use could be reduced by approximately 4.4 billion gallons per year. Resultant direct energy savings are estimated to be 41 million MWh electricity and 240 billion cubic feet of natural gas, with associated CO2 reductions of about 38.3 million metric tons. Based on national averages, indirect energy savings from reduced water supply and treatment energy needs would be about 9.1 million MWh per year, with carbon emissions reductions of 5.6 million metric tons.

- Outdoor water use often drives peak water demands and requires the utilization of marginal water sources with greater energy intensities. Reducing outdoor irrigation—especially during summer months—can result in substantial “upstream” energy savings by reducing water consumption from the most energy-intensive supplies and by avoiding the need to develop additional supplies.

- A 5% reduction in water distribution system leakage would save 270 MGD of water and 313 million kWh of electricity annually, equal to the electricity use of over 31,000 homes. In addition, approximately 225,000 metric tons of CO2 emissions could be avoided.

- If groundwater levels across the United States were to drop an average of 10 feet due to unsustainable water withdrawals, energy demands for agricultural groundwater pumping would increase by approximately 1.1 million MWh per year. Assuming pumping energy is derived from the U.S. electrical grid, associated carbon dioxide emissions would be approximately 680,000 metric tons per year.

- An average sized 1,000 MWh power plant that installs a water reuse system for cooling tower blow-down recovery would reduce the energy demand to produce, distribute and treat water by a net 15%, or enough to power over 350 homes for a year.

- If LID techniques were applied in southern California and the San Francisco Bay area, between 40,400 MG and 72,700 MG per year in additional water supplies would become available by 2020. The creation of these local water supplies would result in electricity savings.
of up to 637 million kWh per year and annual carbon emissions reductions would amount to approximately 202,000 metric tons by offsetting the need for inter-basin transfers and desalinated seawater.

The link between water and energy presents the climate change community with a valuable opportunity to better manage two of our most valuable resources. As the U.S. struggles to reduce its carbon emissions in response to global warming, investments in water conservation, efficiency, reuse and LID are among the largest and most cost-effective energy and carbon reduction strategies available. Furthermore, water is perhaps the most vital ecosystem service that our natural environment provides. As the inevitable impacts of climate change become evident, our freshwater resources and the ecosystems they support will become respectively less reliable and resilient. Smart water policies allow us to mitigate the worst aspects of global warming today, while the consequent improvements in water quantity and river health will provide a critical buffer as humanity and nature adapt to the climate of tomorrow.
Climate change and growing demands already strain our energy and water supplies. It has been projected that under a “business as usual” scenario, electricity demand in the United States (U.S.) will increase by 53% between 2003 and 2030. Much of the country is currently experiencing water shortages, with many of the fastest growing regions in the nation already withdrawing up to five times more water than is naturally replenished through precipitation. Meanwhile, the Intergovernmental Panel on Climate Change predicts that global warming will result in less reliable water supplies, while the efforts to develop lower carbon energy sources could drive a shift toward a more water-intensive energy portfolio. Given these trends, it is imperative that policies at all levels ensure the sustainable management of both water and energy.

The “water-energy nexus” is a broad label for the set of interactions caused when humans develop and use water and energy. The nexus manifests itself in many ways, revealing substantial tradeoffs and opportunity costs associated with the ways we use water and energy. Producing thermoelectric power, for example, requires large amounts of water for cooling, while nearly every stage of the water use cycle involves energy inputs. A better understanding of the water-energy nexus will allow integrated resource planning that optimizes the use of invaluable and increasingly scarce resources.

Energy production in the U.S. requires more water than any other sector. According to the U.S. Geological Survey, 48% of water withdrawals in the United States are used for thermoelectric power production. In addition, water is used for growing biofuels or in the extraction of coal, petroleum and natural gas. To illustrate this connection, consider that a hundred-watt light bulb turned on in drought-stricken Atlanta, Georgia for 10 hours results in the consumption of 1.65 gallons of water (with a carbon footprint of 1.4 pounds).

On the other hand, water use in the U.S. requires significant amounts of energy. Water is heavy at 8.34 pounds to the gallon and energy is required whenever it is moved, treated, heated or pressurized. For many communities, the energy required for supplying and treating water and wastewater constitutes the largest municipal energy cost.
In California, for instance, water-related energy use in 2001 was estimated at 48 million MWh (or 48 thousand GWh) of electricity, plus 4.3 billion Therms of natural gas and 88 million gallons of diesel fuel. This energy use results in approximately 38.8 million metric tons of carbon dioxide emissions annually. Water-related electricity alone accounts for 19% of California’s electricity consumption, while natural gas use—primarily for water heating—accounts for 30% of the state’s natural gas demand. The carbon emissions embedded in California’s water as a result of these energy demands is equivalent to the carbon emissions of 7.1 million passenger vehicles, and would require approximately 9 million acres of pine forest to offset California’s water-related carbon footprint.

Unless our water supplies are properly managed, the carbon footprint of water use in the United States will continue to grow at a time when climate change necessitates reducing carbon emissions. With so many interconnections, what can we safely say is the “carbon footprint” of water use in the United States today? Furthermore, what policies or techniques are available to reduce water-related carbon emissions?

In order to answer these questions, River Network conducted a literature review of primary and secondary research on water use and its associated energy requirements in the United States. This report builds on River Network’s initial estimate of nationwide water-related energy demands by utilizing updated sources and new considerations. To quantify water-related energy use in the U.S., we explored three key research areas:

1. The extent of water-withdrawals across the country by sector,
2. The range of energy intensities for water supply & treatment, and
3. Current estimates of energy in end uses of water.

In Section Four of this report we propose a new base estimate of U.S. water-related energy use and carbon emissions. After establishing the magnitude of water-related energy consumption, we conclude the report by exploring the carbon-reducing potential of various water conservation, efficiency, reuse and low-impact development programs.
Every five years, the United States Geological Survey (USGS) collects data on the nation’s water withdrawals and compiles it in an authoritative report titled *Estimated Use of Water in the United States*. The most recent USGS report on water use contains data collected in the year 2000 and is used for most of this report. (As of 3/31/09 the 2005 report has not been released.)

The USGS defines water withdrawals as “water removed from a ground- or surface-water source for use.” This broad definition refers to all human uses of water, regardless of whether or not the water is returned to the environment or available for later use. Water consumption—or consumptive uses of water—refers to, “that part of water withdrawn that is evaporated, transpired by plants, incorporated into products or crops, consumed by humans or livestock, or otherwise removed from the immediate water environment.” Differentiating between water consumed and water withdrawn is critical to understanding how much water is available for environmental and human uses, and hence necessary for water supply planning.

It should be noted that definitions of terms relating to water use are not always clear and aggregating water use figures from different reports can be misleading. Water may have been measured before or after it was delivered to end users. In many instances it is not metered at all. Return flows may be diverted by another user or returned to the environment to replenish groundwater. The terms “diverted,” “withdrawn” or “consumed” may mean different things to different agencies. Even where water rights are carefully managed under specific beneficial use statues, conveyance losses may not be fully measured.

The way that water use is broken into sectors can also be confusing. Aside from public supplies, nationwide water use data is frequently categorized by end-user. Private end-users are broken down by economic sector (irrigation, industrial, thermoelectric power, mining, aquaculture and livestock) and
“domestic use” (referring to self-supplied households). Therefore, to determine total national water withdrawals by end-use, the public water supplies must also be broken down by end-user.

Many reports do not differentiate between public and private supplies. The Pacific Institute, a well-known research institution focusing on water issues, typically categorizes water users as either urban or agricultural. In this case, urban use refers to the residential, commercial, institutional and industrial sectors, while agricultural uses include irrigating food, fodder and fiber crops. Both urban and agricultural water use can be either public or private, although a large portion of agricultural water is self-supplied. These complications become evident when compared to USGS findings. While agriculture composes the vast majority of the irrigation sector referred to by USGS, uses likely considered urban such as, “Irrigation of golf courses, parks, nurseries, turf farms, cemeteries, and other self-supplied landscape-watering uses also are included.”

The USGS estimates that water withdrawals in the entire United States amount to approximately 408 billion gallons of water per day (GPD) or 149 trillion gallons per year (see figure 1.1). The vast majority of these water withdrawals come from freshwater and surface sources, representing 85% and 79% of total withdrawals, respectively. By sector, thermoelectric power generation accounts for 48% of all water withdrawals and irrigation accounts for 34%—making them the two largest water using sectors. Public water supplies rank third representing 11% of the total.

### Table 1.1 – Estimated Use of Water in the United States by Sector, 2000 (USGS)

<table>
<thead>
<tr>
<th>Sector</th>
<th>Daily Water Use (MGD)</th>
<th>Annual Water Use (MG)</th>
<th>% of Total</th>
</tr>
</thead>
<tbody>
<tr>
<td>Public Supply</td>
<td>43,300</td>
<td>15,804,500</td>
<td>11.00%</td>
</tr>
<tr>
<td>Self-Supply Domestic</td>
<td>3,590</td>
<td>1,310,350</td>
<td>&lt;1%</td>
</tr>
<tr>
<td>Industrial</td>
<td>19,700</td>
<td>7,190,500</td>
<td>5.00%</td>
</tr>
<tr>
<td>Mining</td>
<td>3,490</td>
<td>1,273,850</td>
<td>&lt;1%</td>
</tr>
<tr>
<td>Irrigation</td>
<td>137,000</td>
<td>50,005,000</td>
<td>34.00%</td>
</tr>
<tr>
<td>Livestock</td>
<td>1,760</td>
<td>642,400</td>
<td>&lt;1%</td>
</tr>
<tr>
<td>Aquaculture</td>
<td>3,700</td>
<td>1,350,500</td>
<td>&lt;1%</td>
</tr>
<tr>
<td>Thermoelectric</td>
<td>195,000</td>
<td>711,750,000</td>
<td>48.00%</td>
</tr>
<tr>
<td>U.S. Total:</td>
<td>407,540</td>
<td>148,752,100</td>
<td>100.00%</td>
</tr>
</tbody>
</table>

### Public Supplies

Public water supplies are defined as systems serving at least 25 people through a minimum of 15 connections. Approximately 242 million people receive water from public supplies, representing 85% percent of the population and accounting for 43.3 billion gallons of withdrawals per day. Public supply represents 11% of total water withdrawals and 13% of freshwater withdrawals. Slightly over a third (37%) of all water withdrawn for public supplies came from groundwater sources, with surface water making up the balance. No information on alternative public water supplies—such as desalination or recycled water—was included in the USGS survey.
Because the vast majority of water users receive their supply through public systems, we are particularly interested in where it comes from and how it is used. The USGS did not include data on deliveries in public supply systems for 2000, so information had to be gleaned from the 1995 survey. Approximately 56% of all water that made its way into public systems was delivered to domestic users, with commercial use ranking a distant second, composing 17% of 1995 public demand. Public use and losses accounted for 15%, industrial demand was 12% and thermoelectric power ranked lowest, representing less than 1% of public water demand. Therefore, residential users account for more water demand on public supplies than all other sectors receiving public water combined. Public use and lost water is technically unaccounted for and represents 15% of all public water demands, a staggering volume that should be better tracked in order to minimize lost water.

Conclusions

- Our nation withdraws an estimated 149 trillion gallons per year. Public water systems withdraw 43 billion gallons of water each year and serve 242 million people, or eighty-five percent of the population.
- Residential users acquire more water from public supplies than all other sectors combined.
- Public use and lost water is unaccounted for and represents 15% of all public water demands, a staggering volume that should be better tracked in order to minimize lost water.
- Future research on the water-energy nexus would benefit from a national agreement on how best to measure water withdrawals (water diverted, used, consumed and/or replenished) and consistent definitions of the sectors being measured by end user and water source.
Section Two
The Energy Intensity of Water

The energy intensity of water use (also called virtual or embedded/embodied energy) is the total amount of energy, calculated on a whole-system basis, required for the use of a given amount of water in a specific location. This calculation can vary considerably based on a number of factors. Among the most important aspects are the type and quality of source water, the pumping requirements to deliver water to end-users, the efficiency of the water system and the energy embedded by specific consumer end uses.

Energy intensity values are typically expressed in kilowatt hours because electricity is the predominant energy type for municipal water supply and wastewater treatment systems. While energy sources other than electricity are occasionally used for water supply and treatment, 93% of water providers and 86% of wastewater treatment plants respectively receive 90% and 80% of their operating energy from electricity.

The energy inputs of a typical water-use cycle can be broken down into five basic stages (Figure 2.1):

![Figure 2.1: From Wolff et al., 2](image)

The 5 stages above can often be broken down into additional components. Figure 2.2 depicts a schematic designed by the California Energy Commission and based on work by Dr. Robert Wilkinson. This schematic provides a slightly more detailed look at the different energy inputs in a typical water use cycle. (Turquoise blue represents sources of water, water supplies are shown in light blue, water and wastewater treatment are shown in purple, and end use is shown in beige.) End-use energy is
embedded by the consumer and is the only component not considered in the energy intensity of water supply and treatment. The reuse of wastewater represents an additional component that is found in a growing number of water systems.

The energy intensity of each component of the water cycle can differ considerably, resulting in a wide variability of embedded energy values between water systems. Including wastewater treatment but not including end-use, the energy intensity of municipal water supplies on a whole system basis can range from a low of 1,050 kWh/MG to a hypothetical high upwards of 36,200 kWh/MG (See Table 2.1). For most utilities, energy use varies from 1,250 kWh/MG to 6,500 kWh/MG.20

Table 2.1 – Range of Energy Intensities for Water Use Cycle Segments21

<table>
<thead>
<tr>
<th>Water Use Cycle Segments</th>
<th>Range of Energy Intensity (kWh/MG)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Low</td>
</tr>
<tr>
<td>Water Supply and Conveyance</td>
<td>0</td>
</tr>
<tr>
<td>Water Treatment</td>
<td>100</td>
</tr>
<tr>
<td>Water Distribution</td>
<td>250</td>
</tr>
<tr>
<td>Wastewater Collection and Treatment</td>
<td>700</td>
</tr>
<tr>
<td>Wastewater Discharge</td>
<td>0</td>
</tr>
<tr>
<td><strong>Total:</strong></td>
<td><strong>1,050</strong></td>
</tr>
</tbody>
</table>

**Water Supply Factors**

The type, quality and location of a water supply are the primary factors influencing the energy embedded in a water supply system. Other important factors include water lost in the system due to leaks, the efficiency of water pumps and the spatial and topographical characteristics of the distribution system.

In general, the energy required by most utilities for treatment and distribution of potable water differs from 250 kWh/MG to 3,500 kWh/MG.22 It often takes a great deal of energy to move water,
and pumping costs are directly related to the elevation water must be lifted. Depending on pumping efficiency, between 40 and 80 kWh are required to lift one million gallons of water 10 feet.23 Energy used for groundwater pumping is typically between 537 kWh and 2,270 kWh per million gallons, depending on pumping depth.24 Although some gravity fed surface sources are located above the service area and require no additional pumping, energy is often needed to pump surface water sources as well. For instance, water delivered to Southern California from the Sacramento-San Joaquin Delta passes 2000 feet over the Tehachapi Mountains and requires 9,200 kWh/MG.25

The vast majority of water supplies come from fresh groundwater or surface sources such as rivers, lakes or streams.26 Other sources of water include desalinated seawater, brackish groundwater and recycled wastewater. Table 2.2 provides some generic estimates of the energy intensity for water supplies.

Table 2.2 – Generic Energy Intensity of Water Supply Types27

<table>
<thead>
<tr>
<th>Source Types</th>
<th>Energy Intensity (kWh/MG)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Surface Water (Gravity Fed)</td>
<td>0</td>
</tr>
<tr>
<td>Groundwater</td>
<td>2000</td>
</tr>
<tr>
<td>Brackish Groundwater</td>
<td>3200</td>
</tr>
<tr>
<td>Desalinated Seawater</td>
<td>13800</td>
</tr>
<tr>
<td>Recycled Water</td>
<td>1100</td>
</tr>
</tbody>
</table>

Many water utilities rely on multiple sources of water, much in the same way an electric utility might get its power from multiple power plants. Different sources of water can be embedded with varying degrees of energy depending on quality, location and type of source. This results in water systems that supply units of water with different values of embedded energy throughout the year. Marginal units of water are most likely to have a higher energy factor than the system as a whole, since the least energy-intensive sources available are generally used to meet base load demands due to their lower costs to supply.

In Portland, Oregon, for instance, the Portland Water Bureau relies on two sources to meet its water demands. The primary source, the Bull Run, consists of gravity fed surface water from a protected watershed and requires just 570 kWh per million gallons. The secondary, or marginal, source consists of groundwater withdrawn at the bureau’s Columbia South Shore Well Field, which has to be pumped 4.5 miles south and 750 feet up for storage. Due primarily to these pumping demands, well field water has an energy intensity of approximately 3,675 kWh per million gallons—about 6.5 times greater than the Bull Run supply.

Despite successful water conservation efforts, peaking water demand and limited supplies in the Bull Run during summer months often forces the bureau to use the well field supply, thus increasing electricity costs. In 2006, for instance, the groundwater supply represented 43% of total electricity requirements despite providing only 14% of that year’s water supply.28 Therefore, reducing the demand of water from the well field will have a greater energy benefit than a similar reduction of Bull Run water. This implies that the bureau could optimize energy savings by aggressively targeting summer water use in its conservation programs.
In many cases, the analogy between water and electric utilities continues into the preference for least cost resources, which are always dispatched before more expensive resources if possible. This fact influences the carbon impact of water because the least cost electric resources for most utilities in this country are high carbon, fossil-based fuels such as coal. As major electricity users, utilities may receive a larger-than-average share of their electricity from the cheaper, dirtier sources supplying power to the local grid. Thus the more electricity embedded in water, the higher the carbon impact.

**Wastewater Treatment Factors**

The energy intensity of wastewater treatment depends on the pumping demands for wastewater collection, as well as the level of treatment and size of facility. For most wastewater treatment plants, energy use ranges between 1,000 kWh/MG and 3,000 kWh/MG, although outliers do exist. The largest energy intensity values are as high as 6,000 kWh/MG, or double the high-end of the typical range.29

While wastewater treatment plants are often sited in order to utilize gravity fed wastewater collection, not all plants are located downhill from consumers and many utilities incur pumping costs to move wastewater to the treatment plant. Pumping wastewater is inherently more inefficient than pumping freshwater because pumps are designed to accommodate solids in the wastewater stream.30

The energy intensity of treating wastewater increases with greater levels of treatment and decreases with scale. Table 2.3 consists of average energy intensity values illustrating the relationship between level of treatment, size of facility and energy intensity.

### Table 2.3- Energy Intensity of Wastewater Treatment by Size and Level of Treatment31

<table>
<thead>
<tr>
<th>Treatment Plant Size (million gallons/day)</th>
<th>Unit Electricity Consumption (kWh/million gallons)</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Trickling Filter</td>
</tr>
<tr>
<td>1 MGD</td>
<td>1,811</td>
</tr>
<tr>
<td>5 MGD</td>
<td>978</td>
</tr>
<tr>
<td>10 MGD</td>
<td>852</td>
</tr>
<tr>
<td>20 MGD</td>
<td>750</td>
</tr>
<tr>
<td>50 MGD</td>
<td>687</td>
</tr>
<tr>
<td>100 MGD</td>
<td>673</td>
</tr>
</tbody>
</table>

**Current Trends**

While the current magnitude of the energy required to supply and treat water and wastewater is large, a number of notable trends are likely to increase the energy intensity for water supply and treatment, thus increasing its carbon footprint. The three major trends are 1) greater reliance on marginal water supplies, 2) development of new energy-intensive supplies and 3) regulatory standards requiring higher levels of drinking water and wastewater treatment.
Growing water demand and decreased reliability of many water resources suggests that more water providers will be forced to rely on marginal water supplies with greater energy and carbon emissions costs. As the example of Portland, OR illustrates, marginal water supplies often require significantly more energy than primary supplies (In Portland’s case, 6.5 times more energy is required to pump water from a marginal source compared to the primary supply). A study commissioned by the Portland Water Bureau in 2002 found that global warming will likely decrease the water available from Portland’s primary source (the Bull Run) during the summer, when water demand is highest. In total, it was estimated that the Water Bureau will be required to supply an additional 1.3 billion gallons of water per year from alternative sources, such as the more energy intensive well field supply. Assuming the 1.3 billion gallons of additional water is provided by the well field supply at an energy intensity of 3,675 kWh/MG, the energy required to supply Portland’s water will increase by approximately 4.8 million kWh per year.

Many water utilities already reach or exceed the capacity of their current water supplies and are looking to develop new water sources. As local supplies become increasingly strained, water utilities are forced to pump groundwater from deeper depths or consider inter-basin water transfers or desalination. Seawater desalination is about seven times more energy intensive than groundwater, while groundwater supplies are about 30% more energy intensive than surface water. In California, a state facing a long-term drought coupled with a growing population, about 20 different water agencies are considering desalination. If all of the desalination facilities currently proposed in California were built, desalination would represent 6% of California’s year 2000 urban water demand and significantly increase the energy intensity of California’s water supplies.

Santa Fe, New Mexico offers another example of how new water supplies will likely increase the energy intensity of supplying water in the United States. In April 2009, five Eastern New Mexico farmers filed applications to transfer 2 billion gallons of water per year from their farmlands near Fort Sumner to consumers in Santa Fe. If approved, this water would be pumped nearly 150 miles and 4,000 feet in elevation to reach consumers in Santa Fe. To put this lift in context, the State Water Project (SWP) in California currently has the highest lift of any water system in the world, pumping water 2,000 feet over the Tehachapi Mountains to convey water from northern to southern California. The Santa Fe supply requires twice the elevation climb. Assuming a pumping efficiency of 70% (4.48 kWh/MG) and no water lost due to system leaks, the energy intensity of Santa Fe’s proposed water supply would be about 18,000 kWh/MG for pumping alone. If the proposed 2 billion gallons of water annually is actually delivered through this supply, new energy costs would be about 36 million kWh annually with associated CO2 emissions of about 32,400 metric tons per year.

When drinking water and wastewater discharge standards are made more stringent, the energy required for water and wastewater treatment generally goes up. For instance, in 2001 the U.S. EPA began imposing tougher standards on water providers to control microbial contaminants, such as cryptosporidium a parasite commonly found in lakes and rivers. Recently, pharmaceuticals, endocrine disrupting compounds and personal care products have been detected in the drinking supplies of at least 41 million Americans. Removing these contaminants is an energy-intensive process.
and if water regulations and standards become more strict, the energy intensity of treating water in the U.S could increase significantly.\textsuperscript{41}

Tougher standards are also being enforced for wastewater and stormwater treatment. The EPA has recently implemented tougher rules requiring onsite stormwater treatment.\textsuperscript{42} As a result, millions of gallons of water that previously entered waterways as polluted runoff will now require energy as its treated to acceptable discharge levels. Table 2.3 shows how the energy intensity can more than double when switching between trickling filter to advanced wastewater treatment with nitrification. If tougher standards are adopted requiring more stringent wastewater treatment, the energy intensity of wastewater treatment should increase accordingly.

Conclusions

- The energy intensity of municipal water supplies on a whole system basis can range from a low of 1,050 kWh/MG to a hypothetical high upwards of 36,200 kWh/MG, while a more typical range between 1,250 kWh/MG and 6,500 kWh/MG is found for most water systems. Thus, the energy embedded in the water delivered by public utilities varies widely between systems and within a single system. The wide range of energy intensities suggests that the energy intensity should be determined for specific water systems in order to accurately assess the energy embedded in a community’s water supply.

- The energy intensity of treating wastewater increases with greater levels of treatment and decreases with scale. A typical range for wastewater treatment and collection varies from 1,000 kWh/MG and 3,000 kWh/MG, with some utilities reporting energy intensities as high as 6,000 kWh/MG.

- Current trends indicate that the energy intensity of water supply and treatment in the United States will likely increase given shifts toward a greater reliance on marginal water supplies, the development of new energy-intensive supplies and regulatory standards requiring higher levels of drinking water and wastewater treatment.
Section Three
Estimating Energy in Water End-Uses

Once a water supply reaches a consumer, additional energy is often used to heat, cool, pressurize or purify the water in preparation for its intended use.\(^43\) Energy from sources other than electricity is often embedded in water at end-use, most notably natural gas for water heating. Compared to the other five stages of the water use cycle, end use has the greatest potential for water and energy savings because it saves energy both “upstream,” and “downstream.” Upstream refers to all of the energy required to bring the water to its point of use, while downstream refers to the energy expended to treat and dispose of water.\(^44\)

Energy associated with end-uses of water can be characterized by three typical types: heating, additional pumping and energy used in conjunction with water use that is not directly embedded in water (See Table 3.1).

Table 3.1- Types of Energy Embedded in Water at End-Use\(^45\)

<table>
<thead>
<tr>
<th><strong>Heating</strong></th>
<th>Baths or showers, washing hands, dishes and clothes, industrial processes</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Additional Pumping</strong></td>
<td>Cooling towers, recirculation hot water loops, car washes or high pressure spraying, pressurization for high rise buildings, irrigation pressurization or lifts from canals on farms</td>
</tr>
<tr>
<td><strong>Indirect</strong></td>
<td>Energy used to run an air conditioning compressors that are water cooled</td>
</tr>
</tbody>
</table>

For this discussion, it is important to determine the energy intensity of different end-uses. The Pacific Institute and the NRDC began developing energy intensities and their findings for commercial end-uses can be found in Table 3.2. As shown, energy intensities range between 0 kWh/MG and 207,800 kWh/MG for the commercial end-uses analyzed. Industrial water uses for chilling, process water use, and plant cleaning are also significant and should be explored. Due to the limited number of end-uses analyzed, the range of energy intensities for commercial and industrial end-uses is likely greater than the range shown in Table 3.2.
Table 3.2 - Estimated Energy Intensity of Commercial End-Use

<table>
<thead>
<tr>
<th>Water Use Category</th>
<th>Energy Intensity (kWh/MG)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Kitchen Dishwashers</td>
<td>83,500</td>
</tr>
<tr>
<td>Prerinse nozzles</td>
<td>21,000</td>
</tr>
<tr>
<td>Laundries</td>
<td>35,800</td>
</tr>
<tr>
<td>Water-cooled Chillers</td>
<td>207,800</td>
</tr>
<tr>
<td>Single Pass Cooling</td>
<td>0</td>
</tr>
<tr>
<td>Landscape Irrigation</td>
<td>0</td>
</tr>
</tbody>
</table>

Not every gallon of water conserved by a consumer has the same energy impact. River Network has estimated that end-use energy for residential water use ranges between 0 kWh/MG (for outdoor irrigation or toilet flushing) to 203,600 kWh/MG (for dishwashers). This considers only water heating and might be higher if other energy inputs are considered. We first gathered data on the percentage of hot water typically used for different residential end-uses. From there, we applied the percent of hot water for each end-use to the energy required to heat a unit of water, which was assumed at 0.2036 kWh per gallon based on the energy required to heat water from 55 º to 130 º F ( 75 º F) with an electric water heater. Table 3.3 shows the energy intensities for common residential end-uses.

Table 3.3- Estimated Hot Water Requirements and Energy Intensity of Residential End-Use

<table>
<thead>
<tr>
<th>Water Use Category</th>
<th>Hot Water (%)</th>
<th>Energy Intensity (kWh/MG)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Bath</td>
<td>78.2%</td>
<td>159,215</td>
</tr>
<tr>
<td>Clothes Washers</td>
<td>27.8%</td>
<td>56,600</td>
</tr>
<tr>
<td>Dishwasher</td>
<td>100%</td>
<td>203,600</td>
</tr>
<tr>
<td>Faucet</td>
<td>72.7%</td>
<td>148,017</td>
</tr>
<tr>
<td>Leaks</td>
<td>26.8%</td>
<td>54,565</td>
</tr>
<tr>
<td>Shower</td>
<td>73.1%</td>
<td>148,832</td>
</tr>
<tr>
<td>Toilet</td>
<td>0%</td>
<td>0</td>
</tr>
<tr>
<td>Landscape Irrigation</td>
<td>0%</td>
<td>0</td>
</tr>
</tbody>
</table>

These energy intensities are important for understanding and comparing the energy required—and potential savings through conservation—for common end-uses. However, it is difficult to extrapolate this data without detailed information on how much water is used per end-use. In order to come up with a national estimate of energy required for end-uses of water, we had to take a different approach.

We believe that of the three types of energy inputted at end-uses (heating, additional pumping, indirect), water heating represents the largest share. Due to insufficient data on water use and end-use energy inputs, we decided to look at estimates of total energy use for water heating rather than extrapolate figures based on the energy intensities mentioned above.

Total U.S. Energy Use for Water Heating

Data from the Energy Information Administration (EIA), an agency within the U.S. Department of Energy that collects statistics on energy use within the United States, was used to estimate the
energy embedded in residential and commercial water heating. The agency also collects energy use information in the manufacturing and industrial sectors, but data on water heating in these sectors is currently unavailable.

The residential sector consists of single family and multifamily housing units. Ninety-nine percent (109.8 million) of the 111.1 households in the United States rely on four major fuels for water heating: electricity, natural gas, fuel oil and liquefied petroleum gas (LPG).\textsuperscript{49} The two predominant sources of energy for water heating are natural gas and electricity, accounting for 50% and 40% of the energy (in kWh equivalent) used for residential water heating. Table 3.4 shows the energy use for water heating in the residential sector by fuel source, as well as the kWe for each source.

Table 3.4- Residential H2O Heating by Fuel Source, 2005

<table>
<thead>
<tr>
<th>Fuel Source</th>
<th>Annual Energy Use</th>
<th>kWh Equivalent (billion Kwh)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Electricity (billion kWh)</td>
<td>122</td>
<td>122</td>
</tr>
<tr>
<td>Natural Gas (billion cf)</td>
<td>1,368</td>
<td>153</td>
</tr>
<tr>
<td>Fuel Oil (million gallons)</td>
<td>986</td>
<td>13.4</td>
</tr>
<tr>
<td>LPG (million gallons)</td>
<td>1,642</td>
<td>15.8</td>
</tr>
<tr>
<td>Total</td>
<td></td>
<td>304.2</td>
</tr>
</tbody>
</table>

According to the EIA, “Commercial buildings include all buildings in which at least half of the floor space is used for a purpose that is not residential, industrial or agricultural.”\textsuperscript{50} Using this definition, schools, correctional institutions, buildings used for religious worship and other building types not traditionally considered “commercial” are included under this category. The most recent data available on commercial water heating is from 2003; actual energy consumed for commercial water heating in 2005 is likely higher. Table 3.5 shows energy use for water heating in the commercial sector by fuel source, as well as the kWe for each source.

Table 3.5: Commercial H2O Heating by Fuel Source, 2003

<table>
<thead>
<tr>
<th>Fuel Source</th>
<th>Annual Energy Use</th>
<th>kWh Equivalent (billion Kwh)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Electricity (billion kWh)</td>
<td>26</td>
<td>26</td>
</tr>
<tr>
<td>Natural Gas (billion cf)</td>
<td>338</td>
<td>37.8</td>
</tr>
<tr>
<td>Fuel Oil (million gallons)</td>
<td>131</td>
<td>1.8</td>
</tr>
<tr>
<td>District Heating (Trillion btu)</td>
<td>46</td>
<td>13.5</td>
</tr>
<tr>
<td>Total</td>
<td></td>
<td>79.1</td>
</tr>
</tbody>
</table>

To display different energy sources (such as natural gas or fuel oil) in a consistent kWh unit of measurement, it was assumed that the kWh equivalent equals the amount of electricity available for use if the fuel were used in a thermoelectric power plant. The efficiencies of thermal power plants were assumed to be 40% and 37% for natural gas and petroleum-fired power plants, respectively.\textsuperscript{51} Heating fuel oil and LPG were assumed to have the same efficiency as petroleum. Line losses of 7.2% were also taken into account.\textsuperscript{52} District heating as an energy source for commercial water heating is recorded in Btu’s by the EIA. Because the specific fuel used for district heating was unspecified, a direct conversion to kWh was conducted at a rate of 3,412 Btu/kWh.
Energy embedded in water at end-uses typically represents the largest energy input in the water use cycle. In California, for example, residential, industrial and commercial end-uses of water account for an estimated 58% of the state’s water-related electricity consumption, not counting the additional energy consumed through other fuels such as natural gas and diesel.\textsuperscript{53} Even in San Diego—where water deliveries through the State Water Project and the Colorado River Aqueduct result in a relatively high energy intensity of 6,260 kWh/MG for conveyance—end-use still makes up 57% of the city’s water-related energy consumption.\textsuperscript{54} It is likely, given the sizeable energy requirements for California’s unique system of moving water across the state, that end-use makes up an even larger share of water-related energy consumption in the rest of the country.

While residential water use may be similar from house to house, commercial and industrial uses are not. The mixture of business types and processes makes it hard to find accurate data on water-related energy use in the CII sectors. Information exists in many forms, the most complete covers the State in California, but has not been compiled nationally.

Conclusions

- The energy intensity of different end-uses of water varies drastically with some use requiring no additional energy (e.g. irrigation, toilet flushing) and others requiring up to 203,600 kWh/MG (e.g. dishwasher). Therefore, some water conservation measures will achieve significantly greater end-use energy savings than others.

- While the prospects for reducing energy through water-saving end use strategies may be quite high, national data is scarce.

- Energy embedded in end-uses includes 304 million MWhe for residential water heating, and 79.1 million Mwhe for commercial water heating. These numbers are what River Network will use as base line estimates of national water-related end-use energy consumption.
Section Four
A New Estimate of National Water-Related Energy Use

In the spring of 2008, River Network estimated water-related energy use in the United States by combining data from a 2002 Electric Power Research Institute (EPRI) report on water supply and treatment with statistics on residential water heating in 2001 from the Energy Information Administration (EIA). This calculation was intended to provide a conservative estimate that could be used as for our efforts to raise awareness of the issue until more information became available. At that time River Network concluded that water-related energy consumption in the United States was equivalent to at least 360 million MWh, or 9% of total U.S. electricity demand. No quantification of the carbon emissions associated with water-related energy use was attempted at that time.

To determine the energy required for water supply and treatment, we relied on the findings from Water and Sustainability (Volume 4): U.S. Electricity Consumption for Water Supply & Treatment—The Next Half Century, a report published by EPRI in 2002. This report sought to quantify the energy required for water supply and treatment in the United States in 2000, and provided projections of energy use for each water-using sector through 2050. The projections for 2005 were used in our analysis, however, given the wide variability of energy intensities presented in Sections II and III, there is reason to believe that the EPRI findings represent an unreliable, if not diminutive, estimate of the energy required to supply and treat water in the United States. Despite its potential shortcomings, the EPRI report offers the only available estimate of the aggregate electricity demands of water supply and treatment in the U.S.

Because the source data gathered from the EPRI study was derived from projections based on statistics compiled in 2000, we believe new research should be conducted to verify the precision of EPRI’s findings. A new analysis that disaggregates energy use by source would also be useful, particularly for
understanding the air emissions resulting from water supply and treatment. An assessment of the timing of peak energy use for water supplies on a national and regional basis could provide some useful insight for policy makers and resource planners attempting to integrate water and energy policies. An analysis of the methodology and assumptions used in the EPRI report is provided as Appendix A.

Having conducted a broader review, it is now clear that our initial analysis significantly underestimated the magnitude of water-related energy use in the U.S. By applying updated statistics from the EIA on residential and commercial water heating to the same methodology used in 2008, River Network is now proposing a more accurate baseline estimate 50% greater than our initial findings.

As of this date, a fully comprehensive national analysis of the energy demands associated with water supply, treatment and end-uses has yet to be conducted. This is due to a variety of reasons, including a general lack of awareness of the water-energy nexus, the difficulty obtaining detailed data from utilities and a lack of coordination between researchers and agencies looking at water and energy issues.

River Network’s current estimate of 2005 water-related energy use and associated carbon emissions by sector (derived from EIA and EPRI data) is available in Table 4.1. We believe our estimate provides a baseline estimate of water-related energy use in the U.S. and shows what sectors are responsible for the greatest energy demands. The intent of our analysis is to illustrate the magnitude of water-related energy use by providing a minimum value until a more comprehensive analysis is conducted on the energy embedded in water.

Table 4.1- U.S. Annual Water-Related Energy Use and Carbon Emissions, 2005

<table>
<thead>
<tr>
<th>Sector (“P” = Private supply)</th>
<th>Energy Consumption (Million kWh)</th>
<th>Carbon Emissions (Metric Tons)</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Water Supply and Treatment</strong></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Public Water Supply</td>
<td>31,910</td>
<td>19,681,451</td>
</tr>
<tr>
<td>Public Wastewater Treatment</td>
<td>24,512</td>
<td>15,118,512</td>
</tr>
<tr>
<td>Domestic Supply (P)</td>
<td>930</td>
<td>573,605</td>
</tr>
<tr>
<td>Wastewater Treatment (P)</td>
<td>49,025</td>
<td>30,237,642</td>
</tr>
<tr>
<td>Commercial Supply (P)</td>
<td>499</td>
<td>307,773</td>
</tr>
<tr>
<td>Industrial Supply (P)</td>
<td>3,793</td>
<td>2,339,447</td>
</tr>
<tr>
<td>Mining Supply (P)</td>
<td>509</td>
<td>313,941</td>
</tr>
<tr>
<td>Irrigation Supply (P)</td>
<td>25,639</td>
<td>15,813,624</td>
</tr>
<tr>
<td>Livestock Supply (P)</td>
<td>1,047</td>
<td>645,769</td>
</tr>
<tr>
<td><strong>Subtotal for supply and treatment:</strong></td>
<td>137,864</td>
<td>85,031,764</td>
</tr>
<tr>
<td><strong>End Use (Water Heating)</strong></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Residential</td>
<td>304,200</td>
<td>169,140,000</td>
</tr>
<tr>
<td>Commercial/Institutional</td>
<td>79,100</td>
<td>35,760,000</td>
</tr>
<tr>
<td><strong>Subtotal for End Use:</strong></td>
<td>383,300</td>
<td>204,900,000</td>
</tr>
<tr>
<td><strong>U.S. Total:</strong></td>
<td>521,164</td>
<td>289,931,764</td>
</tr>
</tbody>
</table>
Our findings represent a baseline estimate—not an accurate quantification—of water-related energy use in the United States. Although a much more detailed analysis is needed for a complete understanding of the energy demands of our water use, this effort has yielded a number of useful conclusions. We can now confidently say that nationwide water-related energy use is, at a minimum, equivalent to at 521 million MWh per year, or about 13% of the country’s 2007 electricity consumption.59

The proportion of water-related energy made up by end-uses in our national estimate is higher than values indicated in previous studies such as NRDC’s analysis of San Diego, where energy embedded at end-use accounted for 57% of total water-related energy use.60 The large amount of energy embedded in water implies the potential exists for significant energy and carbon emission reductions through water-oriented strategies.

**Figure 4.1 - U.S. Water Related Energy Use (Chart design by River Network)**

Figure 4.1 shows the annual U.S. water-related energy use for 2005 by sector. To determine the energy used for supplying specific sectors with public water, we used water-use data from USGS *Estimated Water Use in the United States* in 1995 and allocated the energy to each sector in proportion to their respective shares of public water use. It was assumed that public water was allocated to the following sectors: 56% to domestic users, 17% to commercial users, 15% to public use and losses, 12% to industrial users.

**Carbon Emissions**

We determined the carbon dioxide emissions embedded in U.S. water supplies for each of the categories used in our estimate of water-related energy use (Table 4.1 shows the mass of carbon dioxide emissions for water use in 2005). See Figure 4.2 for a breakdown of carbon dioxide emissions by water-use sector.
To calculate carbon emissions, the total amount of energy used in each sector was multiplied by a carbon intensity factor specific to the energy source. It was assumed that all of the energy demands for water supply and treatment were met with electricity. U.S. EPA eGRID data from 2007 (version 1.1) provided the carbon intensity factor for the national electric grid. The carbon intensity assumed for each energy source is shown in table 4.2.

### Table 4.3- Carbon Intensity of Energy Sources

<table>
<thead>
<tr>
<th>Energy Source</th>
<th>Carbon Intensity (in pounds)</th>
</tr>
</thead>
<tbody>
<tr>
<td>U.S. Electric Grid (per kWh)</td>
<td>1.36</td>
</tr>
<tr>
<td>Natural Gas (per cubic foot)</td>
<td>0.12</td>
</tr>
<tr>
<td>Fuel Oil (per gallon)</td>
<td>22.384</td>
</tr>
<tr>
<td>Liquefied Petroleum Gas (per gallon)</td>
<td>12.669</td>
</tr>
</tbody>
</table>

In 2005 CO2 emissions were approximately 6 billion metric tons. We estimate the carbon emissions related to water in 2005 were approximately 290 million metric tons, or 5% of all carbon emissions in the U.S. Water-related CO2 emissions are equivalent to the annual greenhouse gas emissions of 53 million passenger vehicles, or the annual electricity use of over 40 million homes.

### Conclusions

- In 2005 the annual water-related energy use in the United States was equivalent to at least 521 million MWh or 13% of 2007 electricity consumption.

- Residential water heating comprises the largest share of water-related carbon emissions.

- Water-related energy consumption is responsible for approximately 290 million metric tons of carbon dioxide emissions annually. This represents about 5% of the U.S. CO2 emissions in 2005.

- The energy embedded at end-use from water heating alone accounts for 74% of total water-related energy use. Because water heating was the only energy input considered in our estimate, end-use likely represents an even greater proportion of water-related energy use.
Section Five
Saving Energy by Saving Water

This section will explore the potential of water efficiency, reuse and low impact development as energy saving measures.

“Water is one of the few sectors in California’s economy where the same policies can serve both preventative and adaptive global climate change goals. Making more efficient use of water will reduce our demands on water resources and shrink the energy consumption associated with water conveyance, pumping, heating and treatment. California water policies can therefore help the State to adapt to the effects of climate change while also minimizing GHG emissions.”

— California Air Resources Board

As noted by the California Air Resources Board in the quote above, water presents one of the few opportunities to employ strategies that will allow us to simultaneously mitigate and adapt to global warming. Per capita water withdrawals in the United States are among the highest in the world, amounting to 1,430 gallons per day when all sectors are considered. River Network now estimates that energy consumption related to this water use requires the equivalent of at least 521 million MWh of electricity. This figure is likely to grow as communities use up local, low-energy water supplies and are forced to supply water from greater distances or from nontraditional supplies. Unless water demand is curtailed, the energy and carbon emissions embedded in water will continue to grow, at the detriment of our climate and riparian resources.

In 2006, River Network proposed a national goal: that conservation, efficiency, reuse and low impact development could reduce municipal water use on a per-capita basis by 40% over 20 years. We still believe it possible to develop a concerted national program to accomplish that goal. The energy and carbon saved would be highly dependent on the energy-intensity of new water sources being developed which are almost always higher and at a greater environmental cost than those of existing water supplies.
The following strategies would contribute to a national water-related energy reduction program:

1 Conservation and Efficiency

Every drop of water conserved reduces energy consumption and associated carbon emissions, although the exact amount of savings varies. As discussed in Sections II and III, the energy embedded in a given unit of water can vary drastically depending on the water system and the type of end-use. In 2005, the California Energy Commission found that investments in water conservation and efficiency improvements could yield 95% of the energy savings as traditional energy-efficiency programs at 58% of the cost.65

An analysis conducted by the Pacific Institute and the NRDC found that satisfying all growth in water demand with conservation would reduce the energy intensity of the California’s water use by 13%.66 In a separate report on water efficiency in the state, the Pacific Institute concludes that with today’s technology, California could reduce urban water use by about 34% across all sectors.67 As shown in Table 5.1, demand for residential, commercial, institutional and industrial water in California could be diminished by up to nearly 40% in each sector.68

<table>
<thead>
<tr>
<th>Urban Water Use by Sector</th>
<th>Potential to Reduce Use</th>
</tr>
</thead>
<tbody>
<tr>
<td>Residential Indoor</td>
<td>39%</td>
</tr>
<tr>
<td>Residential Outdoor</td>
<td>25% - 40%</td>
</tr>
<tr>
<td>Commercial/Institutional</td>
<td>39%</td>
</tr>
<tr>
<td>Industrial</td>
<td>39%</td>
</tr>
<tr>
<td>Water System (unaccounted-for water)</td>
<td>10%53</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td><strong>34%</strong></td>
</tr>
</tbody>
</table>

If we assume that similar water use reductions are achievable in each of these sectors nationwide, then the potential for water efficiency to reduce water-related energy demands in the United States is large. Where peak water use coincides with peak electric use, the water utility will pay higher costs for that electricity. This gives water utilities financial incentives to reduce these coincident peaks—especially for sources where significant energy is embedded.

A. Residential Indoor

Indoor residential water use is relatively homogenous across the United States. Toilets, clothes washers, showers and faucets account for more than 80% of indoor water use for a typical single family home.70 As such, the majority of indoor water conservation efforts rightfully focus on these end-uses. Overall, per capita indoor water use can be reduced by at least 35% with more efficient water using fixtures and appliances. This translates to annual savings of approximately 35,000 gallons of water for a family of four.71
A study of nearly 100 homes found that prior to retrofitting households with the water efficient fixtures and appliances listed above, only 45% of homes surveyed used less than 150 gallons per day. After the retrofit, the number of houses using less than 150 gallons per day nearly doubled to 88%. Overall, it this study found that retrofitting toilets, clothes washers, kitchen and bathroom sink aerators, toilets and showerheads using existing technology can achieve indoor water savings of 39%, comparable to the magnitude of savings in California estimated by the Pacific Institute.

All indoor water saved in the residential sector results in energy savings from avoided water deliveries and wastewater treatment. Based on national averages, the EPA estimates that if just 1% of American homes replaced their older, inefficient toilets with WaterSense labeled models, the country would save more than 38 million kWh of electricity—enough to supply more than 43,000 households electricity for one month. Furthermore, if every household in the U.S. replaced their major water using fixtures and appliances, the indirect energy savings due to water efficiency would amount to about 9.1 million MWh per year, with carbon emissions reductions of 5.6 million metric tons.

Consumers also directly save energy by reducing hot water consumption. The average reduction in hot water use for households installing efficient fixtures and appliances is 10.8 gallons per day, or a reduction of approximately 20%. If every household in the United States achieved similar savings through water efficiency, residential hot water use would be reduced by approximately 4.4 billion gallons per year. Resultant energy savings are estimated to be 41 million MWh electricity and 240 billion cubic feet of natural gas, with associated CO2 reductions of about 38.3 million metric tons.

**Outdoor**

In the United States, outdoor water use is an estimated 7.8 billion gallons per day, with residential outdoor use averaging 31.7 gallons per capita per day. A typical suburban lawn requires 10,000 gallons of applied water per year and 80 to 90% of outdoor residential water is used for watering lawns, plants and gardens. The amount of water used outdoors is highly dependent on climate and landscape design, ranging between 10 to 75% of total residential water demand. Outdoor water use can be reduced through improved irrigation techniques, weather or sensor based irrigation controllers and water-efficient landscape designs. For instance, Xeriscaping, a landscaping technique developed by Denver Water in 1982, has been shown to achieve water use reductions of at least 50% when compared to traditional landscapes. Even simple fixes, such as using an automatic shutoff nozzle on a hand-held hose can reduce outdoor water use by 5 to 10%

For most applications, outdoor water use does not require any additional energy inputs. Some exceptions include pressure washing devices, ornamental water features, swimming pools or hot tubs. For instance, many large fountains pump as much as 4,000 gallons per hour with lifts between 15 and 23 feet. Assuming one of these fountains continuously pumps 4,000 gallons of water per hour 20 feet high with a pumping efficiency of 65%, approximately 3,400 kWh per year are embedded in the water circulating through the fountain. A fountain consuming this much energy creates 2.4 million metric tons of CO2 emissions annually. While outdoor uses with additional energy inputs present significant opportunities for energy and water savings, they represent a small fraction of total outdoor water use.
Reducing water used for outdoor uses and landscape irrigation will offset the “upstream” energy required to deliver that water to the customer. In the summer months, outdoor water use drives peak demands 1.5 to 3 times higher than a typical winter day. Many water supply systems, such as Portland, OR (See Section 2), are forced to use more energy intensive sources to meet these marginal demands. Therefore, water saved for outdoor irrigation often reduces peak demand and the energy needed to deliver the most energy-intensive water supplies. Because outdoor water use constitutes a large portion of residential water consumption and is typically used during periods when utilities rely on marginal water supplies, the national water and energy savings achievable through outdoor water conservation are likely very significant.

**B. Commercial, Industrial and Institutional**

The commercial, industrial and institutional (CII) sectors use approximately 36,690 million gallons per day and represent between 20 to 40% of billed urban water demand. Potential water savings from efficiency and other conservation measures is typically between 15 to 30% for most communities, with savings as high as 50% possible. While more information on end-uses of water in the CII sector is needed to better understand the direct energy resulting from water efficiency, examples exist showing the potential. For instance, pre-rinse spray valves have been installed in nearly 17,000 restaurants in California. Each valve annually saves approximately 50,000 gallons of water and avoids over 7,600 kWh of electricity or 330 Therms of natural gas, depending on water heater type.

**C. Water Supply and Treatment Systems**

Reducing leaks within a water supply system has the potential for significant energy savings. The actual energy savings achieved by reducing leaks will depend on the overall energy intensity of the system and how far down the water supply chain the leak occurs. Embedded energy accumulates as water moves down the supply chain. For instance, water saved at the local distribution stage will embody the energy of all previous stages, including treatment and conveyance. But water saved during conveyance will not have been embodied with the energy of later steps.

A generally accepted estimate for water lost due to supply system leakage is estimated to be on the order of 10% of total supply, or 5.48 billion gallons daily. It is believed that an aggressive national program aimed at reducing system loss could achieve a 5% reduction in leaks, equal to 0.5% of total water supply. This effort would save 270 MGD of water and 313 million kWh of electricity annually, equal to the electricity use of over 31,000 homes. In addition, approximately 225,000 metric tons of CO2 emissions could be avoided.

**D. Agricultural**

The potential of significant water and energy savings in the agricultural sector certainly exists. The USGS estimates that groundwater withdrawals for irrigation in 2000 were approximately 58 billion gallons per day, or over 21 trillion gallons per year. Many regions across the country withdraw more water out of aquifers than is naturally replenished, causing groundwater levels to drop. As a result, declining aquifers require agricultural users to dig deeper wells and use additional energy for lifting water a greater elevation.
In an area south of the Canadian River in New Mexico, groundwater levels in the Ogalla Aquifer declined 26 feet between 1980 and 1999. If groundwater levels across the United States were to drop an average of 10 feet, energy demands for agricultural water use alone would increase by approximately 1.1 million MWh per year. Assuming pumping energy is derived from the U.S. electrical grid at a carbon intensity of 1.36 pounds CO2 per kWh, associated carbon dioxide emissions would be approximately 680,000 metric tons per year.

While water in the agricultural sector can be saved through site selection, soil amendments, crop rotation and conservation-oriented pricing, at this time we only explore the effects of improving irrigation methods. The three primary types of irrigation systems are flood, sprinkler and drip, with average efficiencies of 73%, 78% and 89% respectively.

Flood irrigation is often gravity fed, with certain systems using additional energy to lift water prior to flooding. Ground water pumping is a major energy input to agricultural water use. Sprinklers and drip irrigation require pressurization which in turn adds to embedded energy. Drip irrigation, the most efficient method of irrigation, adds 632 kWh/MG of embedded energy while flood irrigation typically adds no more than 92 kWh/MG to the energy already embedded from the water supply system.

Although drip irrigation is nearly seven times more energy intensive than flood irrigation, it provides significant water savings. Drip irrigation can result in a net energy savings depending on the source water and quantity considered. On-farm testing of pumps and conducting necessary repairs or improvements can often increase pumping efficiency by 5%-15%. According to the California Energy Commission, “These measures can more than offset the new energy requirements that most often accompany drip system installations.”

Table 5.2 Approximate On-Farm Energy Requirements of Different Irrigation Methods

<table>
<thead>
<tr>
<th>Activity</th>
<th>Approximate Energy Requirements (kWh/MG)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Flood irrigation without on-farm lift</td>
<td>0</td>
</tr>
<tr>
<td>Lifting water 10 feet for flood irrigation</td>
<td>92</td>
</tr>
<tr>
<td>Booster pumping for drip/micro-irrigation</td>
<td>632</td>
</tr>
<tr>
<td>Booster pumping for standard sprinklers</td>
<td>872</td>
</tr>
</tbody>
</table>

2 Water Reuse

The U.S. EPA defines water recycling as “reusing treated wastewater for beneficial purposes such as agricultural and landscape irrigation, industrial processes, toilet flushing, and replenishing a ground water basin.” The two primary types of wastewater are known as greywater and blackwater. Blackwater is commonly known as sewage and is what people generally refer to as “wastewater.” Greywater refers to wastewater that contains fewer concentrations of organic waste than water used for toilets or kitchen sinks, but is nonetheless considered non-potable. In this report, rainwater harvesting is considered an LID technique.
Untreated greywater collected onsite can serve many of the same applications as treated wastewater. In addition to nonpotable applications, wastewater and greywater can be treated up to potable standards and used to replace “virgin” water for any conceivable use, as is being done in Singapore and Orange County, California. Overall, the potential to reduce water-related energy consumption through wastewater and greywater reuse is highly dependent on the energy intensity of specific water and wastewater systems and the level of treatment needed for an intended use.

The “yuck” factor—or the psychological aversion many people have to the concept of reusing wastewater—is a significant barrier to the adoption of widespread wastewater reuse. Furthermore, due to the health and safety concerns around greywater reuse, most states currently ban the practice.

### A. Wastewater

Reused wastewater is often treated to standards meeting or exceeding drinking water requirements and is suitable for a number of applications.\(^98\) The byproduct of existing secondary and tertiary wastewater treatment processes can be substituted for freshwater in a number of applications without additional treatment. Even if additional treatment is required, reusing wastewater can offer a net energy benefit by offsetting more energy-intensive sources.

A study by General Electric found that an average sized 1,000 MWh power plant that installs a water reuse system for cooling tower blow-down recovery would reduce the energy demand to produce, distribute and treat water by a net 15%, or enough to power over 350 homes for a year.\(^99\) The actual net energy benefit of wastewater reuse depends on the level of additional treatment needed (if any at all), the pumping costs of distributing the treated wastewater and the energy intensity of existing freshwater sources.\(^100\)

### B. Greywater

Common sources of greywater include clothes washers, baths, showers and bathroom sinks. Typical uses of greywater include toilet flushing and outdoor irrigation of inedible plants. These are two of the largest uses of water in the typical single-family home,\(^101\) implying that well designed greywater reuse applications can significantly offset municipal water demands and the associated energy for supply and treatment.

While greywater reuse appears to be growing, especially in drought-stricken regions, widespread adoption of greywater reuse does face significant barriers. Due to the presence of potentially hazardous organic matter in greywater, it is not suitable for all uses. Collecting greywater can also be difficult as many household are not designed to properly store and treat greywater. Some strategies do exist for reusing greywater and growing number of products designed for capturing and reusing greywater are hitting the market. A few models provide onsite greywater treatment, making greywater suitable for more applications. Greywater reuse in commercial, institutional and industrial applications should be explored further.
Greywater can also be used to reduce energy use through an increasingly popular method called drain water heat recovery. This method uses a heat exchanger to recover heat from the hot water used in showers, bathtubs, sinks, dishwashers, and clothes washers. The energy used for water heating can be reduced by 30% or more with these devices, with an expected payback range from 2.5 to 7 years.102

3 Low Impact Development

Low Impact Development (LID) refers to comprehensive land planning and engineering design approaches that seek to maintain or enhance the pre-development hydrologic regime of urban and developing watersheds.103 In other words, LID is a stormwater management approach and set of practices that are designed to reduce runoff and pollutant loadings by managing stormwater as close to its source as possible.104 Green roofs, rainwater harvesting, bioretention areas (or rain gardens, bioswales), permeable pavement, and riparian habitat protection are among the most commonly used LID strategies.

LID strategies can reduce the energy required for stormwater treatment, avoid the carbon emissions associated with building traditional infrastructure, reduce aquifer drawdown and provide a “new” local water supply through aquifer storage or rainwater harvesting. While the full extent of energy savings attainable through LID techniques is currently unknown, we explore the potential for energy and carbon emissions reductions using specific examples below.

A study conducted by the Trust for Public Land and American Water Works Association found that 50 to 55 percent of a utility’s treatment costs can be explained by the percentage of forest cover in the source area. The study further concluded that for every 10 percent increase in forest cover, treatment and chemical costs decreased by approximately 20 percent.105 It is unclear precisely how much of these savings are attributable to energy reductions. Given that electricity constitutes between 25 and 40 percent of a typical wastewater treatment plant’s budget and 80 percent of the costs of processing and distributing drinking water,106 one can conclude that the energy savings associated with protecting source water and reducing the contaminant load of stormwater will be significant.

Both rainwater harvesting and aquifer recharge using LID techniques such as bioretention areas have the potential to make available large quantities of water that would otherwise go unutilized. Rainwater can be stored onsite using a simple rain barrel or a larger cistern. Harvested rainwater can be applied directly for outdoor irrigation or treated for a variety of potable uses. The full potential of rainwater harvesting as a water supply has not be quantified, however, a number of case studies exist that illustrate its effectiveness. For instance, Honda of America built a system to capture rainwater for use in the cooling towers of its Marysville Auto Plant in Ohio. The seven-acre, two-pond facility can store 22 million gallons of rainwater and has helped the facility reduce its groundwater usage by 40 million gallons a year.107

A study conducted by the NRDC took into account detailed land use analyses, water supply patterns and information on the energy consumption of local water utilities to determine the water, energy and
carbon emissions reductions achievable through LID techniques in California. The study looked at limited portions of the San Francisco Bay Area and urbanized areas of southern California to conclude that if LID techniques were applied in just these areas, between 40,400 MG and 72,700 MG per year in additional water supplies would become available by 2020. The creation of these local water supplies would result in electricity savings of up to 637 million kWh per year. Based on the carbon intensity of California’s current electricity grid, the annual carbon emissions reductions would amount to approximately 202,000 metric tons. The report’s findings are believed to be conservative. According to its authors, “Far greater water and electricity savings—and associated reductions in greenhouse gas emissions—would additionally result from full application of LID practices statewide.

Aquifer recharge through LID techniques also has the potential to maintain groundwater levels, thus avoiding additional pumping demands that result when groundwater levels drop. Depending on pumping efficiency, between 40 and 80 kWh are required to lift one million gallons of water 10 feet. Utilizing LID to maintain aquifer levels could significantly reduce the energy required for pumping groundwater, especially in regions where groundwater represents the majority of water supplies.

Finally, if a project’s entire lifecycle is considered, LID has the potential to avoid significant greenhouse gas emissions by avoiding a share of the construction costs associated with building traditional water infrastructure. The use of concrete and other materials with a relatively large carbon footprint can be minimized with onsite stormwater containment. Since LID approaches use plants, they have the potential to absorb carbon emissions over their lifecycle, while traditional infrastructure can increase impervious surfaces and increase the energy for treating water.

Conclusions

- If every household in the United States installed efficient fixtures and appliances, residential hot water use would be reduced by approximately 4.4 billion gallons per year. Resultant energy savings are estimated to be 41 million MWh electricity and 240 billion cubic feet of natural gas, with associated CO2 reductions of about 38.3 million metric tons. Based on national averages, indirect energy savings from residential indoor water efficiency is about 9.1 million MWh per year, with carbon emissions reductions of 5.6 million metric tons.

- Outdoor water use often drives peak water demands and requires the utilization of marginal water sources with greater energy intensities. Reducing outdoor irrigation—especially during summer months—can result in substantial “upstream” energy savings by reducing water consumption from the most energy-intensive supplies.

- Potential water savings from efficiency and other conservation measures in the CII sector is typically between 15 to 30% for most communities, with savings as high as 50% possible.

- A 5% reduction in water distribution system leakage would save 270 MGD of water and 313 million kWh of electricity annually, equal to the electricity use of over 31,000 homes. In addition, approximately 225,000 metric tons of CO2 emissions could be avoided.
If groundwater levels across the United States were to drop an average of 10 feet due to overuse, energy demands for agricultural groundwater pumping would increase by approximately 1.1 million MWh per year. Assuming pumping energy is derived from the U.S. electrical grid, associated carbon dioxide emissions would be approximately 680,000 metric tons per year.

An average sized 1,000 MWh power plant that installs a water reuse system for cooling tower blow-down recovery would reduce the energy demand to produce, distribute and treat water by a net 15%, or enough to power over 350 homes for a year.

If LID techniques were applied in southern California and the San Francisco Bay area, between 40,400 MG and 72,700 MG per year in additional water supplies would become available by 2020. The creation of these local water supplies would result in electricity savings of up to 637 million kWh per year and annual carbon emissions reductions would amount to approximately 202,000 metric tons.
The magnitude of water-related energy use in the U.S. is considerable. At 521 million MWh, water-related energy use is equivalent of 13% of U.S. electricity consumption and has a carbon footprint of at least 290 million metric tons. Substantially more water and energy use data is needed before national, regional and local decision-makers can gain a comprehensive understanding of the energy embedded in the nation’s water supplies. Despite this lack of data, a plethora of water management and water policy options currently exist that could significantly reduce energy and carbon emissions. We recommend the following actions be included in a broad effort to reduce the energy and carbon emissions associated with water use in the United States:

- Explore ways to integrate water and energy policies at the federal, state and local levels to ensure the sustainable management of both resources.
- Develop a standard methodology for water utilities to quantify the energy intensity of their water supplies and benchmark their energy usage.
- Create national guidelines for reporting water use across end-use sectors.
- Introduce a national requirement for all water and wastewater service providers to regularly report on annual energy consumption and the energy intensities of their respective water supplies.
- Allow public access to energy intensity values so that consumers are aware of the energy and carbon implications of their water use.
- Conduct research on the energy embedded during end-uses of water, particularly commercial and industrial uses. A better understanding of the energy required for different end-uses would allow water conservation programs to target the most energy-intensive uses of water and optimize carbon emissions reductions.
- Launch pilot water conservation, efficiency, reuse and Low Impact Development programs that measure the energy savings achieved and can serve as case studies.
• Educate the public about the relationship between water and energy so that consumers can make informed decisions about their water use.

The link between water and energy presents the climate change community with a valuable opportunity to better manage two of our most valuable resources. As the U.S. struggles to reduce its carbon emissions in response to global warming, investments in water conservation, efficiency, reuse and LID are among the largest and most cost-effective energy and carbon reduction strategies available. Furthermore, water is perhaps the most vital ecosystem service that our natural environment provides. As the inevitable impacts of climate change become evident, our freshwater resources and the ecosystems they support will become less reliable and resilient. Smart water policies allow us to mitigate the worst aspects of global warming today, while the consequent improvements in water quantity and river health will provide a critical buffer as humanity and nature adapt to the climate of tomorrow.
The Carbon Footprint of Water

13 Hutson et al, 11

14 Ibid, 13


17 Ibid, 5

18 Carlson, Steven W. and Adam Walburger. Energy Index Development for Benchmarking Water and Wastewater Utilities, United States: AWWA Research Foundation; New York State Energy Research and Development Authority; California Energy Commission, 2007. (Pgs. 21 & 73)

19 Klein et al, 7

20 Carlson, 14 & 68

21 Table derived from Klein, page 9. With modified low values for wastewater collection and treatment and water distribution. Modified values assume mostly gravity fed distribution and gravity wastewater collection with 50 MGD trickling filter treatment facility.

22 Carlson, 14

23 Assumes optimum pumping efficiency of 75% (4.2 kWh/MG/1ft lift) and low efficiency of 40% (7.9 kWh/MG/1ft lift), from The University of California Cooperative Extension, Tulare County Available at: <http://cetulare.ucdavis.edu/pubgrape/ig696.htm>.


26 Hutson et al, 4

27 From Pacific Institute’s “Water to Air Model,” factors do not include distribution. These values were developed by the Pacific Institute to serve as default values for their Urban Water to Air model, which can be accessed at: <http://www.pacinst.org/resources/water_to_air_models/index.htm>

28 Energy intensity values based on electricity consumption and water production data from the Portland Water Bureau (Personal email from Rod Allen 8/11/08). The values we used represent an energy intensity based on the aggregate water production and electricity consumption from 2004-2007. The methodology applied in Portland did not follow the one developed by Robert Wilkinson and applied in California. This was due to the unavailability of facility-specific data and may result in a low estimate of the energy intensity of groundwater. The data that used for the energy intensity of groundwater did not include electricity use for distribution pump stations and other facilities, while our assessment of surface water took into account these additional energy costs (Personal email from Rod Allen, 9/16/08). As a result, the difference in energy intensity between Bull Run and CSSWF water is likely greater than our conclusions suggest.

29 Carlson, 68

30 Klein et al, 38


34 EPRI, 2-3


36 Cooley et al, 2.


38 Wolff et al, 2

39 Assumes electricity from NM grid at 1.99 lbsCO2/MG, supra note 5.


43 Wolff et al., 18

44 Ibid, 18

45 Ibid, 33

46 Ibid, 33


48 Energy intensity of residential end use was estimated by applying the percent of hot water to the energy required to heat a unit of water, which was assumed at 0.2036 kWh per gallon based on the energy required to heat water from 55 ° to 130 ° F (△ 75 ° F) with an electric water heater.


52 U.S. Climate Technology Program

53 Klein et al, 9

54 Wolff et al., 33

55 Carbon emissions resulting from water supply and treatment assume all energy comes from electricity with a carbon intensity of 1.36 lbs CO2/kWh, based on EPA eGRID 2007 Version 1.1.
EPRI, 1-5.

EIA 2005, Table WH3


Wolff et al., 32

Supra note 8.

Supra note 5.


Klein et al, 128

Wolff et al., 34

Gleick et al., 2

It should be noted that Pacific Institute’s findings are specific only to California and it is unclear how they relate nationally. It could be assumed, however, that extrapolating California’s figures nationally may underestimate the potential of water efficiency in the United States. California enacted legislation requiring 1.6 gallon per flush toilets and other water efficient fixtures in January 1992, preceding the enforcement of similar federal requirements by two years. Furthermore, the state is known for its longstanding and aggressive resource conservation and environmental protection policies, including incentives for water and energy conservation that are not available in many other states. The market for water efficient devices is likely more saturated in California than in other states, hence the potential to reduce water use in the United States as a whole is likely greater than the 34% estimated in California.

ICF, 7-23


Ibid, 27


Assumes water savings of 68 gallons per household per day (Aquacraft, 3) and 111.1 million households (EIA). Energy intensity is estimated at 1,500 kWh/MG for water supply and 1,800 kWh/MG for wastewater treatment (EPRI). Carbon emissions based on national electric grid, 1.36 lbs/kWh.

Ibid, 5


Assuming 54% of households use natural gas and 46% use electric water heating based on EIA “Water-Heating Energy Consumption in U.S. Households by Type of Housing Unit, 2001.” Available at: <http://www.eia.doe.gov/emeu/recs/recs2001/ce_pdf/waterheat/ce4-4c_housingunits2001.pdf>
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Notes

78 Vickers, 141
79 Ibid, 147
80 Ibid, 222
81 Greenhouse Gas Equivalencies Calculator, EPA
82 Vickers, 140
83 Ibid, 232
84 Ibid, 235
86 ICF, 7-23
87 Ibid, 7-28
88 Greenhouse Gas Equivalencies Calculator, EPA
89 Hutson et al, 20.
92 Pumping is assumed to be 60% efficient (5.25 kWh/MG). Agricultural pumping efficiency is assumed to be less than that for public water systems due to the smaller scale of agricultural systems and the likelihood agricultural that pumps are not optimized or replaced as frequently.
93 Cooley, More with Less, Pg. 36
94 Klein et al, 45
95 Cooley, More with Less, Pg. 44
98 Klein et al, 28
99 Bolze
100 Klein et al, 28
103 Low Impact Development Center, Inc. Available at: <http://www.lowimpactdevelopment.org/index.html>.
Notes


109 Assumes carbon intensity of California electric grid is 0.7 lbs CO2 per kWh based on EPA eGRID2007 Version 1.1

110 Horner, 1

111 Supra, note 19

112 Pumping is assumed to be 60% efficient (5.25 kWh/MG). Agricultural pumping efficiency is assumed to be less than that for public water systems due to the smaller scale of agricultural systems and the likelihood agricultural that pumps are not optimized or replaced as frequently.

113 Bolze

114 EPRI, 1-1


116 Public supplies go to a variety of end-use sectors, including residential, commercial, industrial, etc.
Appendix ~
Issues with EPRI Assumptions

In 2002, the Electric Power Research Institute (EPRI) published a study on the electricity required to supply and treat water in the United States. The report looked at public water agencies, publicly and privately owned wastewater treatment facilities and self-supplied water to answer the following question: “Will there be sufficient electricity available to satisfy the country’s need for fresh water?”

To this end, the report concluded that about 4% of the country’s electricity is devoted to water supply and treatment. Electricity was the only energy source analyzed, and there was no attempt to develop a detailed assessment of the total energy demands of water supply and treatment. In short, the EPRI study was not intended to be a definitive report on the subject. Despite the limited scope of the EPRI study, it nonetheless provides the best available assessment of the energy requirements of water supply and treatment in the United States.

Approach
The EPRI study based its analysis on publicly available, secondary sources and was completely transparent about the methodology employed. EPRI categorized sectors based on the same characteristics as USGS in its Estimating Water Use in the United States series; that is, between public supplies and private end-use sectors, with the addition of publicly and privately owned wastewater treatment works. To analyze the electricity demands of each sector, EPRI determined the per unit electricity requirements of surface water and groundwater withdrawals in each sector and applied these values to water use information from USGS the U.S. EPA. Projections were carried out based on population growth as characterized by the U.S. Census Bureau. The energy intensity and total projected energy use in 2005 for wastewater treatment and each end-use sector can be found in Table A.1.
Public Supplies

For public water supplies, per unit energy requirements were determined for surface water and groundwater sources using previously published data. These per unit averages were then applied to the U.S. EPA's Safe Drinking Water Information System inventory, which at the time consisted of about 30,000 water systems. Based on size of the water system and source type of water, each water system was assigned an energy factor (kWh/MG) and total electricity use was determined based on the amount of water withdrawn by each public system. The average per-unit energy requirement assigned for surface water supplies is 1,406 kWh/MG, and 1,824 kWh/MG for groundwater. EPRI's projections for electricity consumption in 2005 (the report was written in 2000) assume that increases in water use will correspond directly with the U.S. Department of Commerce Census Bureau projections of population growth.

Table A.1 Energy Intensity and Total Energy Use for Water Supply and Treatment

<table>
<thead>
<tr>
<th>Sector (“P” refers to private supply)</th>
<th>Energy Use, 2005 (million kWh)</th>
<th>Surface Water (kWh/MG)</th>
<th>Groundwater (kWh/MG)</th>
<th>Wastewater (kWh/MG)</th>
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</thead>
<tbody>
<tr>
<td>Public Water Supply</td>
<td>31910</td>
<td>1406</td>
<td>1824</td>
<td>-NA-</td>
</tr>
<tr>
<td>Domestic (P)</td>
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<td>-NA-</td>
<td>700</td>
<td>-NA-</td>
</tr>
<tr>
<td>Commercial (P)</td>
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<td>300</td>
<td>700</td>
<td>2500</td>
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<td>Industrial (P)</td>
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<td>300</td>
<td>750</td>
<td>2500</td>
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<td>Mining (P)</td>
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<td>750</td>
<td>2500</td>
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<td>Livestock (P)</td>
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<td>700</td>
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<td>700</td>
<td>-NA-</td>
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<td>Power Generation (P)</td>
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<td>300</td>
<td>800</td>
<td>-NA-</td>
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<td>Public Wastewater Treatment</td>
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<td>955</td>
<td>1322</td>
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<td>Private Wastewater Treatment</td>
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<td>-NA-</td>
<td>2500</td>
</tr>
</tbody>
</table>

Public Wastewater Treatment Systems

EPRI employed a similar methodology for public wastewater treatment works (POTW’s) as it did for public water supplies. First, per unit energy consumption was determined for different levels of water treatment and size of facilities. These energy factors were then applied to the U.S. EPA's 1996 inventory of POTW’s based on size and level of treatment characteristics. Total energy use for this POTW’s was estimated by extrapolating the treatment volume and energy factors of specific POTW’s in the database. For details on the assumptions made for more complex wastewater treatment processes, see EPRI, 3-6.
It should be noted that the general trend in the U.S. is towards higher treatment standards. Since the energy intensity of wastewater treatment is directly related to the level of treatment, higher treatment standards will likely increase the overall energy demands of wastewater treatment, however, this increase might be offset if larger treatment plants are built and efficiency is improved due to economies of scale. Table 3.1 shows the relative energy intensities for wastewater treatment based on facility size and level of treatment.

**Private Wastewater Treatment Works:**

Privately operated wastewater treatment works refers to the approximately 23,000 privately operated treatment facilities that are typically used onsite with industrial plants and commercial operations. Because detailed statistics on privately operated plants was not available, EPRI assumed a prototypical per unit energy-intensity of 2,500 kWh/MG, due to the significantly smaller scale of these facilities. It was noted that more aggressive wastewater treatment standards would increase unit electricity requirements by 5 to 10 percent over the next 20 years, and that more privately operated plants would be built over this period. Due to the lack of information on privately operated treatment works, when conducting projections, it was assumed that total electricity demand for privately operated plants is double that for publicly owned treatment works.

More research on the energy use of privately operated treatment plants is needed for a better estimate of the energy demands in this sector. The value used by EPRI was selected more or less arbitrarily. If the magnitude of energy required for private wastewater treatment is accurate, then this sector represents the largest share of water-related electricity demand and should be investigated further.

**Other End-Use Sectors**

The methodology EPRI used for all of the remaining end use sectors in this report is similar to that described above. Basically, each sector was assigned a per unit energy consumption factor for surface and groundwater supplies based on respective water use characteristics. Energy intensity values can be found in Table A.1. The per unit energy consumption factors were then applied to USGS data for surface and groundwater withdrawals to determine total annual electricity requirements for each sector.

**Potential Issues**

For reasons described below, it is difficult to ascertain the relative accuracy of EPRI’s projections of the energy requirements for water supply and treatment in the United States. The very fact that we have to rely on projections calculated nearly a decade ago for energy use in 2005 implies that a much better estimate is attainable if a new analysis were conducted. While this EPRI report provides the best available estimate for the electricity demands of water supply and treatment in the U.S., a quick review of the methodology and assumptions used in the EPRI study shows the need for more research in this area.
According to the principal investigator assigned to the report, the goal of the project was to determine if enough electricity would be available to meet our future water needs, not to conduct a definitive analysis of water-related energy use in the United States. The report relied on secondary sources, and there has been no follow up to assess the accuracy of the report’s conclusions. Because the EPRI report looked solely at electricity demands, the full breadth of energy impacts are not necessarily captured in EPRI’s analysis. For instance, wastewater treatment plants rely heavily on the use of natural gas to heat anaerobic digesters.

The methodology used by EPRI seems to imply that the conclusions represent a kWh equivalent of energy demands. However, it is unclear if an alternative methodology that disaggregates energy sources and inputs along the water supply cycle might result in notably different conclusions. Furthermore, aggregating all energy use as electricity makes estimates of carbon emissions and air pollution impacts from water supply and treatment more difficult.

The use of per-unit averages for surface water and groundwater, without taking into account other water sources, presents some potential problems when analyzing public water supplies, particularly when carrying out projections of electricity demands. First, a significant and growing number of public water supplies rely on nontraditional water sources, including recycled wastewater, brackish groundwater, desalinated sea water and imported water. For instance, 15% of all U.S. water withdrawals are saline, typically in the form of brackish groundwater. Brackish supplies are increasingly converted to freshwater in regions where alternative supplies do not exist. As shown in Chapter 2, nontraditional supplies typically require substantially more energy on a per unit basis than the local surface or groundwater sources which water providers have traditionally relied on. There is no indication that EPRI considered the impact of more energy-intensive water sources in their analysis.

In addition, there is reason to believe that a large proportion of new water demands will be met by sources with greater energy intensities. As local supplies become increasingly strained, water utilities are forced to pump groundwater from deeper depths or consider inter-basin water transfers or desalination. If all of the desalination facilities currently proposed in California were built, desalination would represent 6% of California’s year 2000 urban water demand. Seawater desalination is about seven times more energy intensive than groundwater. Water pumping represents about 80-85% of the total electricity consumption for surface water supplies and practically all of the electricity used for groundwater supplies. This suggests that relying on deeper wells, more remote surface supplies and nontraditional water sources will significantly increase energy demands. There is no indication that EPRI considered the trend towards more energy-intensive water sources in their analysis.

On the other hand, one could argue that the EPRI conclusions are overestimating actual electricity use, since electricity requirement estimates include both purchased and self-generated electricity. Self generated electricity is occasionally generated within a water system and can offset electricity use. Examples include hydroelectric generation in gravity fed systems or cogeneration at wastewater treatment plants. According to the EPRI authors, not taking into account self generated electricity
means that “on a percentage basis the projected purchased electricity requirement for water supply and wastewater treatment will be substantially less.”

EPRI did not include projections of electricity requirements for water supplies in the thermoelectric sector. Despite being the largest water using sector, water-related electricity use in the thermoelectric power sector was not forecasted because water use in this sector is expected to decline on an absolute basis.\(^{15}\) This could be a false assumption due to unforeseen trends in electricity production related to addressing climate change. The effects of certain forms of power production (such as carbon capture and sequestration) on water demands for thermoelectric power production are currently unknown, and might actually result in a net increase in water required for power production.\(^{16}\) An updated analysis should consider multiple scenarios for future electricity production and cooling technologies in order to show how trends in the thermoelectric sector will affect the energy required to supply water.

In conclusion, the EPRI report provides an excellent starting place for understanding the magnitude of water-related energy demands, but more research is needed. Because the report was not intended to provide a definitive analysis, a new study with the explicit purpose of quantifying the current and future energy demands of water supply and treatment is long overdue. The EPRI report relied on data that is, in many cases, well over a decade old. As interest grows in this subject, a more detailed examination will be necessary to assess the full extent of water-related energy demands and their associated greenhouse gas emissions.
Appendix Notes ~

1 EPRI, 1-1


3 Public supplies go to a variety of end-use sectors, including residential, commercial, industrial, etc.


5 EPRI, 2-3 – 2-4

6 Ibid

7 Ibid, 3-12

8 Ibid, 3-13

9 Ibid, 1-5

10 Supra, note 89

11 Klein et al, 38

12 Hutson et al, 4

13 Cooley et al, 2.

14 EPRI, 2-2

15 EPRI, 1-6

Acronyms & Abbreviations

CEC – California Energy Commission
CO2 – Carbon Dioxide
CII – Commercial, Industrial and Institutional Sectors
CSSWF – Columbia South Shore Well Field
eGRID – Emissions & Generation Resource Integrated Database
EIA – Energy Information Administration
EPA – Environmental Protection Agency
EPRI – Electric Power Research Institute
GPD – Gallons per day
GPF – Gallons per flush (toilets)
kWh – kilowatt hour
kWh/MG – kilowatt hours per million gallons of water
kWhe – kilowatt hour equivalent
LID – Low Impact Development
LPG – Liquefied petroleum gas
MG – Millions of gallons
MGD – Million gallons daily
NRDC – Natural Resources Defense Council
PI – Pacific Institute
POTW – Publicly-owned wastewater treatment works
REUWS – Residential End-Uses of Water Study
USGS – United States Geological Survey
River Network's mission is to help people understand, protect and restore rivers and their watersheds.

The Carbon Footprint of Water

Bevan Griffiths-Sattenspiel and Wendy Wilson

INNECC1

River Network

01/05/2009

The decisions being made today regarding the management of water and energy resources will profoundly affect our economic health and quality of life. We must ensure that our freshwater resources are managed in a sustainable way so that they are protected for future generations. This paper provides an overview of how water management strategies can protect our freshwater resources while reducing energy and carbon emissions.

Through our analysis of primary and secondary research, we estimate that U.S. water-related energy use is at least 521,000 terawatt-hours per year, of which 85% is for powering water treatment and distribution facilities. This represents an estimated 15% of total U.S. water-related energy use. Water represents 5% of all U.S. carbon emissions and is equivalent to the emissions of over 62 coal fired power plants.1


River Network
National Office
520 SW 6th Avenue, Suite 1130
Portland, OR 97204
Phone: 503-241-3506
Fax: 503-241-9256
info@rivernetwork.org

www.rivernetwork.org